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Delaware Bay New Jersey Oyster Seedbed Monitoring Program 2020 Status Report

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Executive Summary

The 2020 Seedbed Monitoring (SBM) Program tracked oyster size, dermo disease and oyster mortality monthly at six fixed sites, two additional sites of interest and one shellplant site. The Program also continued its long-term disease analyses for the annual Fall Oyster Stock Assessment Survey by assessing meat condition and dermo disease from 22 beds as well as MSX disease data from seven fixed monitoring sites. COVID-19 restrictions delayed sampling until July and precluded sampling intermediate transplant sites and all but the current year shellplant.

Temperature was near seasonal averages for the period sampled. Moderate freshwater inflow throughout the year maintained salinity near seasonal averages for much of the year. Mean oyster size increased on most beds. Dermo disease followed typical seasonal and spatial patterns, but levels were below average overall with highest levels at Shell Rock and Cape Shore such that oysters are entering winter with relatively lighter infections as in 2019. The low prevalence of dermo likely contributed to the relatively low levels of mortality observed during 2020.

Long-term spatial patterns of dermo show a general increase from upper to lower bay beds, but levels in the center of the fishery (Cohansey to Shell Rock) continue to be more prevalent and variable in recent years, and appear to be sustaining consistently higher levels of dermo disease than oysters further down bay. Nevertheless, overall levels remain relatively low with respect to the time series. Long-term mortality also continues to show a general increase from upper to lower beds, excluding periods of freshets. The long-term annual patterns from the Fall survey continue to indicate an attenuation in both duration and amplitude of interannual dermo and mortality cycling. In fact, bay-wide mortality no longer appears to be cycling with dermo and has decreased from 20-30% in the 1990s to less than 20% over the past five years. MSX was detected at low prevalence and intensity on lower Bay beds in Fall 2020, and in other areas of the Bay throughout the year.

The overall picture continues to be one of improvement, but remains highly dependent upon environmental conditions, particularly temperature, salinity and Delaware River discharge in any given year. Increased freshwater inflow, even with freshet driven mortality events, has been beneficial in curtailing dermo related mortality. Continued monitoring of disease and mortality across the natural seedbeds, on transplants and on shell plants is warranted to evaluate performance and to inform management of the resource, particularly in the face of climate change and upstream management of reservoirs that impact freshwater inflow. As production in the lower bay increases via aquaculture and revitalization of leased grounds, monitoring efforts should expand to those areas since pathogens have no allegiance to wild or farmed populations.

Introduction

The Delaware Bay Oyster Seedbed Monitoring (SBM) Program tracks disease, growth and mortality of oysters on the Delaware Bay, New Jersey public oyster beds located in the upper portion of the Bay (Figure 1). The purpose is to provide information that supports the sustainable management of the oyster resource in this region of the bay. Oyster production that occurred on privately owned leases, oyster farms or in waters outside the New Jersey portion of the Delaware Bay oyster fishery is not the focus of this report though some information may be included where relevant.

Oyster mortality on the Delaware Bay oyster beds is caused by a variety of factors including predation, siltation, freshets, disease and fishing. Prior to 1957, predation by oyster drills was a primary concern with their abundance and distribution determined by salinity which was controlled by the amount of freshwater inflow (Carriker 1955). Since the appearance of *Haplosporidium nelsoni* (the agent of MSX disease) in 1957, disease mortality has been the primary concern (Powell et al. 2008). Following a severe and widespread MSX epizootic in 1986, the Delaware Bay population developed significant resistance to MSX disease that extends into low salinity regions where MSX is not typically prevalent in oysters (Ford and Bushek 2012). Nevertheless, routine monitoring continues to detect the MSX parasite in Delaware Bay and naïve oysters quickly succumb to the disease indicating that virulence remains high (Ford et al. 2012). In 1990, an epizootic of dermo disease occurred and changed the population dynamics of the system further. Dermo disease is a form of the molluscan disease perkinsosis that is specific to the eastern oyster *Crassostrea virginica*. It is caused by the alveolate protist *Perkinsus marinus*. Its appearance in 1990 was not the first occurrence of *P. marinus* in Delaware Bay. Prior occurrences were associated with importations of oysters from the lower Chesapeake Bay (Ford 1996) and were often short-lived, dying out once importations ceased and following a cold winter. The 1990 appearance of dermo disease was not associated with any known importations but was related to a regional warming trend after which the documented northern range of *P. marinus* was extended to Maine (Ford 1996). It is likely that *P. marinus* was present in the bay for many years prior to 1990 at levels below detection or at least not causing levels of mortality to warrant concern. Since 1990, dermo disease has been a major source of oyster mortality in Delaware Bay and a primary concern for managing the oyster fishery and the oyster stock (Bushek et al. 2012).

Following the appearance of dermo disease in 1990, average mortality on the seedbeds, as assessed by total box counts during the fall survey, has fallen into three major groups: Low Mortality (LM) beds (formerly called the upper seedbeds), Medium Mortality (MM) beds (formerly called the upper-central seedbeds), and High Mortality (HM) beds (formerly called central and lower seedbeds). These designations are positively correlated to salinity from about 5 to 20 ppt. Higher salinity generally promotes better growth and meat quality but also favors predation and disease. A group of beds above the low mortality region was added to the survey in 2007 after reconnaissance indicated a high abundance of oysters in a region the fishery had exploited in the past and wished to do so again. These beds were collectively designated Hope Creek in 2007, but were subsequently subdivided into Hope Creek, Fishing Creek and Liston Range and categorized as the Very Low Mortality (VLM) beds in reference to the level of disease-induced mortality they experience – the VLM beds experience little disease, but episodic

high mortality occurs in response to freshets (Munroe et al. 2013). Current area management strategies separate Shell Rock (SR) from the original medium mortality region and further subdivide the remaining medium mortality region beds into Medium Mortality Transplant (MMT) and Medium Mortality Market (MMM) beds (Figure 1) based on how they are managed within the fishery. Additional details on management strategies and actions are available in annual stock assessment workshop reports at <http://hsrl.rutgers.edu/SAWreports/index.htm>.

The majority of fresh water entering the system comes from the Delaware River and tributaries located above the oyster beds. Additional inputs from several tributaries that enter the bay adjacent to the seedbeds (Hope Creek, Stow Creek, Cohansey River, Back Creek, Cedar Creek and Nantuxent Creek) combine with the geomorphologic configuration of the shoreline to influence salinity, nutrients, food supply, circulation and flushing in complex ways. These factors undoubtedly interact to influence larval dispersal, recruitment and growth, disease transmission dynamics and, ultimately, disease mortality (Wang et al. 2012).

The temporal and spatial sampling efforts of the Oyster Seedbed Monitoring Program are designed to continually develop a better understanding of factors influencing oyster growth, disease and mortality to inform management and sustain a viable fishery as well as a healthy oyster population and a functional ecosystem. A major objective is to identify seasonal and interannual patterns of disease, mortality, recruitment and growth through time. The core effort monitors six sites along the salinity gradient on monthly basis and a spatially comprehensive survey in the Fall. The monitoring supports additional directed research and sampling efforts that are necessary to develop deeper insights of the dynamics controlling the oyster population within the Delaware Bay ecosystem. As funding permits, these efforts include monitoring transplants (i.e., oysters moved from upper to lower seedbeds), shellplants (i.e., shell placed directly on the seedbeds to increase the supply of clean cultch for recruitment), and replants (i.e., cultch planted in the lower bay high recruitment zone near the Cape Shore then moved and replanted on the seedbeds) as well as other natural events (e.g., freshets) and additional experiments that may be sanctioned. The 2020 objectives for the Oyster Seedbed Monitoring Program were to:

1. Continue the standard monthly time series monitoring New Beds, Bennies, Shell Rock, Cohansey, Arnolds, and Hope Creek for size, mortality and dermo disease
2. Conduct dermo and MSX assays and determine condition indices for each bed sampled during the 2020 Fall Stock Assessment Survey
3. Monitor growth, disease and mortality on 2018 through 2020 shell plantings
4. Monitor growth, mortality and disease on the 2019 and 2020 intermediate transplants

Objectives 1 and 2 comprise the basis of the long-term program that provides fundamental information necessary for both immediate and long-term adaptive management of the resource. These objectives also provide essential baseline/background information against which the success of other objectives and independent research can be evaluated. Objective 1 began in 1998 with five beds (Arnolds, Cohansey, Shell Rock, Bennies and New Beds). In 2010 Hope Creek was added as part of the monthly monitoring program. Objective 3 was initiated as part of the Delaware Bay Oyster Restoration program designed to enhance recruitment on the seedbeds. Shell planting is an annual effort of the management plan for sustaining and

rebuilding the oyster beds, scaled by available funds. Objective 4 examines the performance of the intermediate transplant program that moves oysters downbay from upbay beds. This activity provides access to a portion of the resource that is otherwise unavailable to direct market harvest, but was available to the former “Bay Season” seed fishery (Fegley et al., 2003). In addition to sustaining the industry it helps to rebuild and sustain harvested beds.

Methods

COVID-19 Impacts: The COVID-19 pandemic reached the United States in January 2020 and by mid-March had shutdown many activities across the US. As a result, sampling planned to begin in April was delayed until July and reduced from two to one sampling trip per month. The Shell Fisheries Council voted to cancel any intermediate transplants but did manage to complete one shell plant. With only one sampling trip available, the SBM program elected to focus on the long-term monitoring and forgo monitoring transplants and shellplants from prior years. The lone shell plant was added to the monthly monitoring trip as described below. This strategy allowed the most critical information to be collected.

Monthly monitoring occurred at the six long-term sites along a transect spanning the salinity gradient from Hope Creek to New Beds as well as two additional sites of interest (Nantuxent and Cape Shore). However, sampling began in July due to COVID-19 restrictions; although samples were able to be obtained from Shell Rock and the Cape Shore in June and are presented below. No transplant or shellplant sites were monitored due to limitations from COVID-19 restrictions except the current year shellplant on Bennies Sand which was added to the long-term monitoring trip. Monthly reporting to the Delaware Bay Section of the New Jersey Shell Fisheries Council provided timely information on seasonal changes for management and harvest needs. A spatially comprehensive sampling occurred during the annual Delaware Bay New Jersey oyster stock assessment in Fall 2020. All data were evaluated and compared to prior years to provide insight into inter-annual patterns, long-term trends, and factors affecting the oyster stock.

Figure 1 depicts the sampling locations for the 2020 Annual Fall Oyster Stock Assessment with beds outlined in black and area management regions indicated by blue lines. Management activities and this report reference both regions and beds as appropriate. Beds that fall within the jurisdiction of the state of Delaware comprise about 10-15% of the oyster population in the main stem of the Bay but are not considered in the report nor shown in Figure 1. Details on regions, beds and sampling design are provided in Powell et al. (2008 and 2012) as well as Alcox et al. (2017). Briefly, the beds shown in Figure 1 were divided into grids measuring 0.2 x 0.2 minutes of latitude and longitude (roughly 26 acres or 10.5 hectares each). Monthly samples were collected at fixed stations using a composite bushel of three 1-minute tows with a 0.81 m wide oyster dredge from the R/V James R Joseph. Dots in Figure 1 represent locations of grids selected via a stratified random sampling design for the Fall oyster stock assessment; a subsample of which, generally one high quality and one medium quality, were selected for Fall disease sampling. Quality is determined by relative oyster density within each bed as described in Alcox et al. (2017). When ranked by oyster abundance, high quality grids contain 50% of the total oyster abundance, medium quality the next 48% and low quality grids are rarely sampled as they contain less than 2% of the oyster abundance on a bed.

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Monthly samples were collected from July through November for Objectives 1, 3 and 4 as indicated in Tables 1 and 2. Table 3 identifies beds that have been monitored since 1990 as part of the long-term Fall dermo monitoring program that is affiliated with the Annual Fall Oyster Stock Assessment. Table 4 specifies the grids sampled during the 2020 Annual Fall Oyster Stock Assessment to complete Objective 2.

To complete Objective 1, three one-minute tows with a 0.81 m (2.7 ft) oyster dredge were collected at each site using about 14 m (46 ft) of cable from the R/V James W Joseph. Bottom water temperature and salinity were recorded with a handheld YSI® Pro2030 instrument at each site. A composite bushel (37 L total volume with one third coming from each dredge tow¹) was created and then sorted to enumerate gapers (= dead oysters with meat remaining in the valves), boxes (= hinged oyster valves without any meat remaining) and live oysters. Boxes were further categorized as new (= no indication of fouling with little sedimentation inside valves) or old (= heavily fouled and/or containing sediments) to provide an indication of recent mortality. These data were used to estimate mortality as described by Ford et al. (2006). Up to one hundred randomly selected oysters from the composite bushel were returned to the laboratory where shell heights (hinge to bill) were measured to determine size frequency from each site. Care was taken to avoid any bias in sampling oysters by systematically working through the sample until 100 oysters were identified. It is understood that the sampling gear will bias the collection toward larger animals (Powell et al. 2007), but such bias is presumed constant across sampling dates and countered to some extent by clumping when oysters attach to one another (Morson et al. 2018). Twenty individuals representing the size frequency distribution were then sacrificed for Ray's fluid thioglycollate medium assay (RFTM, Ray 1952, 1966) to determine prevalence and intensity of dermo infections. The percent of oysters in the sample with detectable infections is termed the prevalence. Each infection was then scored (i.e., weighted) for intensity using the "Mackin scale" from zero (= pathogen not detected) to five (= heavily infected) after Ray (1954). These values, including zeros, were averaged to produce a weighted prevalence (WP), which provides an estimate of the average disease level in the sample of oysters (Mackin 1962, Dungan and Bushek 2015). The average intensity of infections, which excludes samples scored as zero, was similarly determined. Though related and similar, each measure provides a different understanding of how disease impacts the population.

Samples for Objective 2 were collected during the Annual Fall Stock Assessment Survey using the commercial oyster boat F/V HW Sockwell. The stock assessment survey consists of a stratified random sampling of the medium and high quality grids on the 23 beds that are outlined in Figure 1 and listed in Table 3 (see Ashton-Alcox et al. 2017 for survey method details). The two lowermost beds, Ledge and Egg Island, contain very few oysters and are sampled in alternate years; Ledge was sampled during 2020. After samples were collected for the stock assessment, the remaining catch was searched to collect oysters for disease analysis, size frequency and condition as indicated in Table 4. Oysters for disease analysis were collected to represent the general size distribution of oysters in the sample, excluding spat. Oysters for size frequency and condition index were collected without regard to size. Dermo was diagnosed as described above. MSX was diagnosed using standard histology (Howard et al. 2004).

¹ At Arnolds and Hope Creek, sample volumes were halved due to small size of the oysters.

Objectives 3 and 4 were not completed due to COVID-19 impacts. After the loss in sales, the industry decided with NJDEP to forgo the intermediate transplant. Due to limitations in staff availability, scheduling conflicts and social distancing requirements, there was insufficient time to schedule a second monthly sampling trip so no samples were collected from prior year shell plants, although we were able to add the Bennies Sand shell plant to the monthly disease sampling trips on October 1 and December 3. On the shellplant site, three to five 1-minute dredge tows were searched on deck for planted shell containing live or dead oysters until 100 live oysters attached to planted shell were collected. All boxes and gapers encountered during this process were collected. If five tows were insufficient to collect 100 oysters the effort was stopped and used what was collected to that point. Care was taken to avoid sampling bias while sorting the catch by working systematically through the sample until 100 live spat or oysters were collected. Boxes were enumerated and categorized as new or old as described above. Live oysters attached to planted shell were returned to the laboratory for size measurements ($n = 50-100$ per site). No disease sampling was performed on the 2020 shellplants as it was in its first year and not expected to have contracted any disease by this point.

Results and Discussion

Freshwater Inflow. The Delaware River Basin Commission is tasked with maintaining sufficient flow to prevent upward movement of the salt line (defined here as $250 \text{ mg/L} = 0.25 \text{ ppt}$) below the city of Philadelphia to maintain drinking water standards, protect industries from corrosive effects of salt water and to protect aquatic life located further downstream (DRBC 2021). This is done by maintaining a minimum flow at Trenton via the metered release of water from reservoirs located in the watershed. Reservoirs are also used to store water for other purposes and as catch basins for flood control. When full, water must be released to be prepared for flood control. River flow during 2020 was closer to median levels with several punctuated pulses of water (Figure 2). This contrasts the more consistent levels of high freshwater flow observed in the preceding two years. Discharge in 2020 more closely followed the seasonal pattern of declining discharge from April to November but was punctuated with large intermittent flows. High discharge decreases water residence time over the oyster beds as it reduces salinity, both of which are associated with reductions in disease and can lead to increased mortality on the uppermost oyster beds (Munroe et al. 2013).

Temperature and Salinity. Temperature and salinity are arguably the most important environmental factors controlling oyster growth, reproduction, disease and mortality. The conditions observed over the seedbeds during 2020 were more or less typical with respect to the past 22 years. Water temperatures measured during 2020 collections followed a typical seasonal cycle with little spatial variability across the seedbeds although sampling did not begin until July (Figure 3A). Spawning temperatures had been reached by the time sampling was initiated. Some warmer than average temperatures were measured on the final sampling at several sites though the last two sites measured were much cooler. The warmer temperatures reflect the generally warmer temperatures that were prevalent in the air, while the two closer to normal corresponded to a change in tide so this discrepancy may be a result of water mass differences. In general, the seasonal temperatures measured were near average levels measured since 1999 with (Figure 4A) but there were warmer waters than normal present at the end of the season. Salinity followed the typical estuarine gradient, increasing from upbay to downbay beds (Figure

3B) but was some what higher on average than the 22-yr mean (Figure 4B) until the end of the season when a large pulse of freshwater inflow (Figure 2) was associated with a rapid drop in salinity.

Oyster size. Shell height (measured hinge to bill) roughly corresponds to age and therefore provides insight into both the size and age structure of the population. Seasonal changes in a population's mean shell height may be affected by growth, recruitment and mortality (both natural mortality and fishing mortality). Mean shell height increased during 2020 on most beds but remained relatively steady on others (Figure 3C). Counterintuitively, average size often decreases over the season as small spat become large enough to be measured while larger older animals are harvested or die. In 2020, the overall average size measured increased, but did not exceed the long-term average (Figure 4C). Figure 5 shows how oyster size has changed annually and shows a cyclical pattern that is likely reflective of the interplay between recruitment and mortality such that mean size increases when mortality and recruitment are low while decreasing as recruitment increases along with mortality of larger sized oysters. Current size frequencies are dominated by smaller oysters with a mean of 62 mm (2.4 inches).

Dermo Disease. Dermo prevalence (the percent of the population with detectable infections), weighted prevalence (WP; the average intensity of dermo in the population, including uninfected oysters) and intensity (the average level of infections in infected animals only) followed typical spatial and seasonal patterns (Figures 3D-F) but were generally lower than long-term means (Figures 4D-F). That is, each measure of dermo disease increased to a peak in fall with levels increasing with the salinity gradient. These observations support a continued suppression of dermo disease that is likely related to elevated levels of freshwater inflow during several previous years as well as the periodic spikes in freshwater flow during 2020 (Figure 2). Additionally, an influx of small oysters on Bennies has continued to reduce the percent of oysters with detectable infections there. This decline in prevalence will consequently reduce the weighted prevalence even without a decrease of the average infection intensity. As a result, the population as a whole entered the winter with a relatively low level of dermo, but some of the important lower bay beds such as Shell Rock and Nantuxent were still relatively heavily infected.

The situation at the Cape Shore was entirely different with high levels of dermo detected throughout most of the monitoring period (Figure 3D-F). Here, a hatchery line produced from wild broodstock collected from the Cape Shore was monitored to provide an index of disease pressure. This stabilizes variation arising from different culture environments and methods such as intertidal vs subtidal, source and age of seed, husbandry differences among farms, and other factors. In 2020, these oysters became heavily infected with dermo in June and sustained heavy infections through the fall. This is partly due to the fact that the oysters monitored at the Cape Shore were from a single cohort of near market- or market-size oysters. Corresponding mortality data is not available, but anecdotal information suggests it was relatively high. Expansion of monitoring onto subtidal areas should be considered as these populations will undoubtedly interact with the fishery via disease transmission and reproduction.

Mortality. The low levels of dermo disease just described was associated with relatively low levels of mortality (Figures 3G-H and 4G-H). Figures 3I and 4I show rates of mortality that are parallel to the long-term mean indicating the same rate of increase. However, the offset is

exaggerated due to the delayed sampling start and the cumulative nature of this graph. However, plots G and H in Figures 3 and 4 indicate that mortality was in fact lower than long term means during the latter half of 2020. An epizootic is defined as a sudden increase in the appearance or intensification of a disease that may or may not be associated with mortality. Under this definition, despite the widespread prevalence and seasonal intensification of dermo disease, Delaware Bay did not experience a dermo epizootic during 2020, but the potential for an epizootic to develop and cause significant mortality remains high.

Transplants, shellplants and replants. As explained above, only one shellplant (Bennies Sand) was executed during 2020 and no former plantings were able to be monitored. Spat were detected on the shellplant on October 1st and averaged 17 mm with no mortality detected. By December size had increased to 20.3 mm with little mortality detected.

Long-Term Fall Patterns. Examination of dermo prevalence, weighted prevalence and mortality by bed indicated a continued significant departure from long-term patterns during 2020 (Figure 6). The long-term patterns typically increase from upper to lower bay beds, but since 2013, dermo prevalence and weighted prevalence have been highest in the central portion of the fishery with the highest levels often on or around Shell Rock. The processes that make this a productive oyster region may similarly make it conducive for dermo disease. Fall 2020, dermo levels were below mean long-term values on many beds, and often below the 95% confidence intervals; but several beds from Ship John to Nantuxent exceeded their long-term mean prevalence. More striking were the low Fall box count fractions on high mortality beds. Those beds saw mortality levels that were uncharacteristically low, often near 50% of the respective long-term means. Mortality was also below long-term means above Shell Rock, with Shell Rock possessing the highest level of mortality measured, albeit not higher than the long-term mean.

Figure 7 depicts annual dermo prevalence, weighted prevalence and box-count estimated mortality from 1989 to 2020 for each mortality region. Each parameter generally decreases from high to low mortality regions. Exceptions are predominantly related to freshets causing mortality in the low mortality regions. Dermo prevalence and weighted prevalence track each other well within and across regions, but mortality patterns on the low and very low mortality regions are distinct from the medium and high mortality regions. Within the high and medium mortality regions, mortality lags disease by about one year. In the LM and VLM regions, mortality is nearly out of phase with dermo disease indicating that dermo is not a primary cause of mortality in these regions. Dermo intensity was much more volatile in the early portion of the time series, but this volatility has dampened in the latter half of the time series (Figure 7B). This dampening also corresponds to a reduction in Fall box count mortality (Figure 7C).

Many factors such as temperature, salinity and recruitment are known to influence dermo disease (Villalba et al. 2004) but the confluence and interaction of these factors is difficult to predict. Moreover, while there is some understanding of how these factors influence spatial and seasonal variation in dermo disease, it is less clear how they interact to influence interannual variation. The bay wide data continue to indicate an attenuation of dermo-induced mortality in the three successive epizootics across the medium and high mortality regions (Figure 8). This observation could be entirely environmentally driven or it could indicate an increase in tolerance (the relative ability of an oyster to survive an infection of a given intensity) or resistance (the

ability of an oyster to limit the development of an infection) to dermo disease. Alternatively, dermo virulence may have declined over time. Lagged correlations between river flow and WP produce a significant negative correlation (Bushek et al. 2012). Additional analyses as well as directed studies and experiments are necessary to develop a better understanding of what factors are at play and, more importantly to this assessment, whether or not management strategies can improve the situation.

Figure 9 depicts the regional mortality rates from each fall assessment since 1990 as a function of dermo disease level (weighted prevalence). Bushek et al. (2012) demonstrated that once weighted prevalence begins to exceed 1.5 mortality begins to increase exponentially. In Figure 9, VLM and LM regions show no increase in mortality with dermo infection level because all infections are below the 1.5 threshold – the high mortality events in the VLM were a result of freshets. A relationship begins to develop across the medium mortality regions as infections increase. This relationship is strongest across the high mortality region where it explains approximately 46% of the annual variability in mortality. The 2020 data points all show relatively low mortality, even on the HM region where dermo levels were approaching 2. Highest mortality was on SR (17%) where dermo WP was 2.5 and had been above 2.0 since August. The pulses of fresh water entering the system appear to be associated with curbing dermo levels sufficiently to eliminate it as a leading source of mortality during 2020. This relationship warrants additional study and coordination with the entities managing water flow through the Delaware River Basin Commission.

Because MSX has not been problematic on the seedbeds for nearly two decades, samples from only seven beds along the upbay-downbay gradient have been examined during the fall survey (Table 4). MSX was detected in just three of the 140 oysters assayed; a prevalence of just 2.1% (Figure 10A). One infection was detected at each of Bennies, New Beds, and Ledge (Figure 10C); all downbay sites in higher salinity where MSX disease is more likely to occur (Ford et al. 2012). All infections were rare (less than 10 plasmodia) to very light (11-100 plasmodia) with no systemic or advanced infections observed (Figure 10C). Hatchery spawned Cape Shore natives were also tested June through November 2020. MSX was detected at 10% prevalence in June, July, and August, and at 5% in September, October and November with a systemic infection observed in September. Previous years have found MSX distributed across the seed beds and these data confirm its continued presence in the bay albeit at low levels. MSX remains a threat to the Delaware Bay oyster population as it continues to cause mortalities elsewhere along the East Coast. Therefore, it remains an important component of the monitoring program to understand sources of mortality from year to year. Because MSX can cause mortality in spring and appears to be more prevalent in the lower bay, it is recommended that some level of routine monitoring of MSX occur throughout the year to improve surveillance.

Science Advice

- Continue to examine the spatial and temporal relationships between environmental drivers of temperature, salinity and fresh water inflow on disease and mortality. Long-term patterns now provide a clear indication that dermo levels drop following freshets resulting in a net positive effect on the population (through reduced mortality). The

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potential of controlling disease and mortality through coordination of reservoir releases up the estuary should be explored with appropriate agencies.

- Because of the complex relationships between prevalence, intensity and weighted prevalence of dermo disease and how they change with temperature and salinity, consider plotting long-term seasonal patterns by bed to look for further insights.
- Investigate the potential evidence for the development of dermo disease resistance and/or attenuation of dermo virulence. Plot the relationship of disease by size class and explore it spatially and temporally for changes.
- Take advantage of freshet-induced mortality and the degradation of a large input of boxes as a mechanism to get a natural estimate of disarticulation rates. It may be possible to do this on lower bay beds during periods of low mortality.
- Consider where and when mortality is occurring during the year to help interpret fall mortality patterns.
- Revisit prior analyses of inshore versus offshore disease and mortality.
- Compile condition index data, although highly variable, to show current year versus long-term means by bed along the bay axis.

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Table 1. 2020 sampling schedule for the NJ Delaware Bay Oyster Seed Bed Long-term Monitoring Program. COVID-19 restrictions prohibited sampling during April, May and June. The six long-term sites are Hope Creek grid 64, Arnolds grid 18, Cohansey grid 44, Shell Rock corner of grids 10, 11, 19 & 20, Bennies grid 110 and New Beds grid 26. Nantuxent grid 10 and the Rutgers Cape Shore Lab were the additional sites of interest that were sampled in 2020. Cape Shore was sampled by foot and dates were not always coincident with other sites but generally within 2 or 3 days. June 2, 2020 samples from Shell Rock were obtained from a commercial harvester. Weather and scheduling conflicts resulting from COVID-19 impacts delayed the mid-September sampling to October 1. Shellplant and transplant sites are described in Table 2. Parameters measured include temperature, salinity, dissolved oxygen, counts of live oysters and boxes, size frequency (shell height), and dermo levels.

Date	Samples	Vessel	Captain
June 2, 2020	Shell Rock and Cape Shore, disease only		
July 9, 2020	6 long-term sites + 2 extra sites	NJDEP RV James W. Joseph	Andrew Hassall
August 18, 2020	6 long-term sites + 2 extra sites	NJDEP RV James W. Joseph	Andrew Hassall
October 1, 2020	6 long-term sites + 2 extra sites and 1 shellplant site	NJDEP RV James W. Joseph	Andrew Hassall
October 19, 2020	6 long-term sites + 1 extra site	NJDEP RV James W. Joseph	Andrew Hassall
December 3, 2020	6 long-term sites + 2 extra sites and 1 shellplant site	NJDEP RV James W. Joseph	Andrew Hassall

Table 2. Additional enhancement sites sampled during 2020. Due to extenuating circumstances related to COVID-19, only the current year shell plant was sampled during 2020. Normally, the number of additional sites requires a separate sampling day, but in 2020 the lone sampling site listed here was added to the sampling trips listed in Table 1.

Bed	Grid	Plant material	Plant yr
Bennies Sand	34/35	ocean quahog	2020

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Table 3. Record of collections for annual fall dermo monitoring since 1990. X indicates bed was sampled in respective year for that column. Prior to 2008, not all beds were sampled. Beginning in 2008, all beds were sampled every year except Ledge and Egg Island which were alternated annually due to a general lack of oysters. Beds are listed approximately by latitude, although some lie at the same latitude with different longitudes.

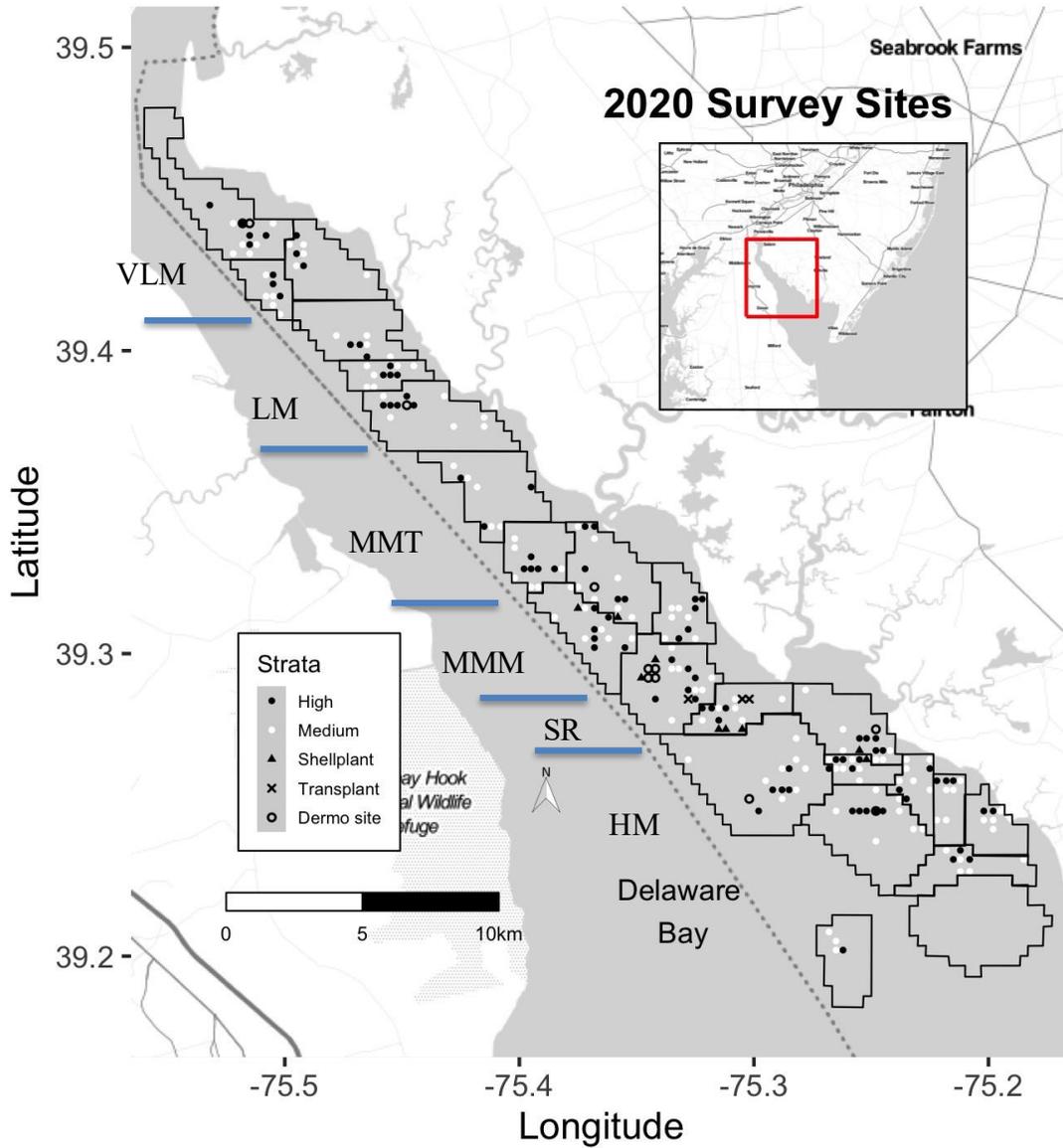
SEEDBED	90	91	92	93	94	95	96	97	98	99	2000	01	02	03	04	05	06	07	08	09	--	--	18	19	20
Hope Creek (HC)																		X	X	X	--	--	X	X	X
Liston Range (LR)																			X	X	--	--	X	X	X
Fishing Creek (FC)																			X	X	--	--	X	X	X
Round Island (RI)	X	X	X	X	X	X	X	X	X	X	X	X	X		X	X	X	X	X	X	--	--	X	X	X
Upper Arnolds (UA)														X	X	X	X	X	X	X	--	--	X	X	X
Amokls (AR)	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Upper Middle (UM)																	X	X	X	X	--	--	X	X	X
Middle (MI)	X	X	X	X	X			X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Cohansey (CO)	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Sea Breeze (SB)														X	X	X	X	X	X	--	--	--	X	X	X
Ship John (SJ)	X	X	X	X	X		X			X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Shell Rock (SR)	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Bennies Sand (BS)	X	X	X	X	X			X	X	X	X	X	X		X	X	X	X	X	X	--	--	X	X	X
Bennies (Ben)	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Nantuxent (Nan)		X		X		X		X		X	X	X			X	X	X	X	X	X	--	--	X	X	X
Hog Shoal (HS)		X		X						X		X	X	X	X	X	X	X	X	X	--	--	X	X	X
New Beds (NB)	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Strawberry (ST))	X		X		X								X	X	X	X	X	X	X	X	--	--	X	X	X
Hawks Nest (HN)	X		X		X		X		X		X		X	X	X	X	X	X	X	X	--	--	X	X	X
Beadons (Bea)	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	X	--	--	X	X	X
Vexton (Vex)										X		X	X	X	X	X	X	X	X	X	--	--	X	X	X
Egg Island (EI)	X	X	X	X	X	X	X	X		X	X	X		X		X		X		X	--	--		X	
Ledge Bed (LB)			X		X				X		X		X		X		X		X		--	--	X		X

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Table 4. 2020 Delaware Bay Oyster Seedbed Stock Assessment Survey grids sampled for dermo, MSX, condition index (CI) and size frequencies. Numbers represent grid ID or the number of oysters processed.

<u>Bed</u>	<u>Grid</u>	<u>Dermo</u>	<u>MSX</u>	<u>CI</u>	<u>Bed</u>	<u>Grid</u>	<u>Dermo</u>	<u>MSX</u>	<u>CI</u>
Hope Creek	43	10		15	Shell Rock	25	10		15
Hope Creek	86	10		15	Shell Rock	51	10		15
Hope Creek	66			10	Shell Rock	36			10
Hope Creek	91			10	Shell Rock	38			10
Hope Creek	63		20	0	Shell Rock	10,11		20	0
Fishing Creek	16	10		15	Bennies Sand	6	10		14
Fishing Creek	25	10		15	Bennies Sand	16	10		15
Fishing Creek	10			12	Bennies Sand	8			11
Fishing Creek	43			8	Bennies Sand	33			10
Liston Range	12	10		15	Bennies	57	10		14
Liston Range	30	10		15	Bennies	102	10		16
Liston Range	17			10	Bennies	84			10
Liston Range	24			10	Bennies	103			10
Round Island	8	10		15	Bennies	110		20	0
Round Island	24	10		15	Nantuxent	20	10		15
Round Island	17			10	Nantuxent	25	10		15
Round Island	18			10	Nantuxent	6			10
Upper Arnolds	5	10		15	Nantuxent	16			10
Upper Arnolds	10	10		15	Hog Shoal	1	10		14
Upper Arnolds	9			10	Hog Shoal	10	10		14
Upper Arnolds	22			10	Hog Shoal	3			11
Amokds	15	10		15	Hog Shoal	14			11
Amokds	45	10		13	New Beds	23	10		14
Amokds	6			10	New Beds	59,60	10		11
Amokds	67			12	New Beds	21			13
Amokds	18		20	0	New Beds	6,17			12
Upper Middle	64	10		15	New Beds	26		20	0
Upper Middle	71	10		15	Strawberry	24,25	10		28
Upper Middle	48			10	Strawberry	10	10		10
Upper Middle	65			10	Strawberry	5			6
Middle	51	10		13	Strawberry	20,28			10
Middle	10	10		14	Hawks Nest	13	10		15
Middle	34			12	Hawks Nest	27	10		15
Middle	38			11	Hawks Nest	26			10
Cohansey	2	10		15	Hawks Nest	3			10
Cohansey	57	10		15	Beadons	4	10		10
Cohansey	4			10	Beadons	9,8	10		28
Cohansey	72			10	Beadons	3,15			9
Cohansey	44		20	0	Vexton	9	10		15
Sea Breeze	19	10		12	Vexton	11	10		10
Sea Breeze	31	10		14	Vexton	4			14
Sea Breeze	16			12	Vexton	10			11
Sea Breeze	36			12	Ledge	7,15	10	10	40
Ship John	16	10		15	Ledge	23	10	10	10
Ship John	32	10		15					
Ship John	28			10					
Ship John	38			10					
Total beds							22	7	22
Total grids							48	10	99
Total oysters							440	140	1097

Figure 1. Footprint of the Delaware Bay, NJ public oyster beds (aka ‘seedbeds’). Black lines demarcate named beds (see Alcox et al. 2017) with management regions approximated by blue lines (abbreviations as in text). The sites for the 2020 stock assessment survey are indicated by dots. A stratified random sampling program identified black and white dots for high and medium density strata, respectively, whereas x denotes transplant sites and triangles were shellplant sites. See Alcox et al. (2017) for full description.



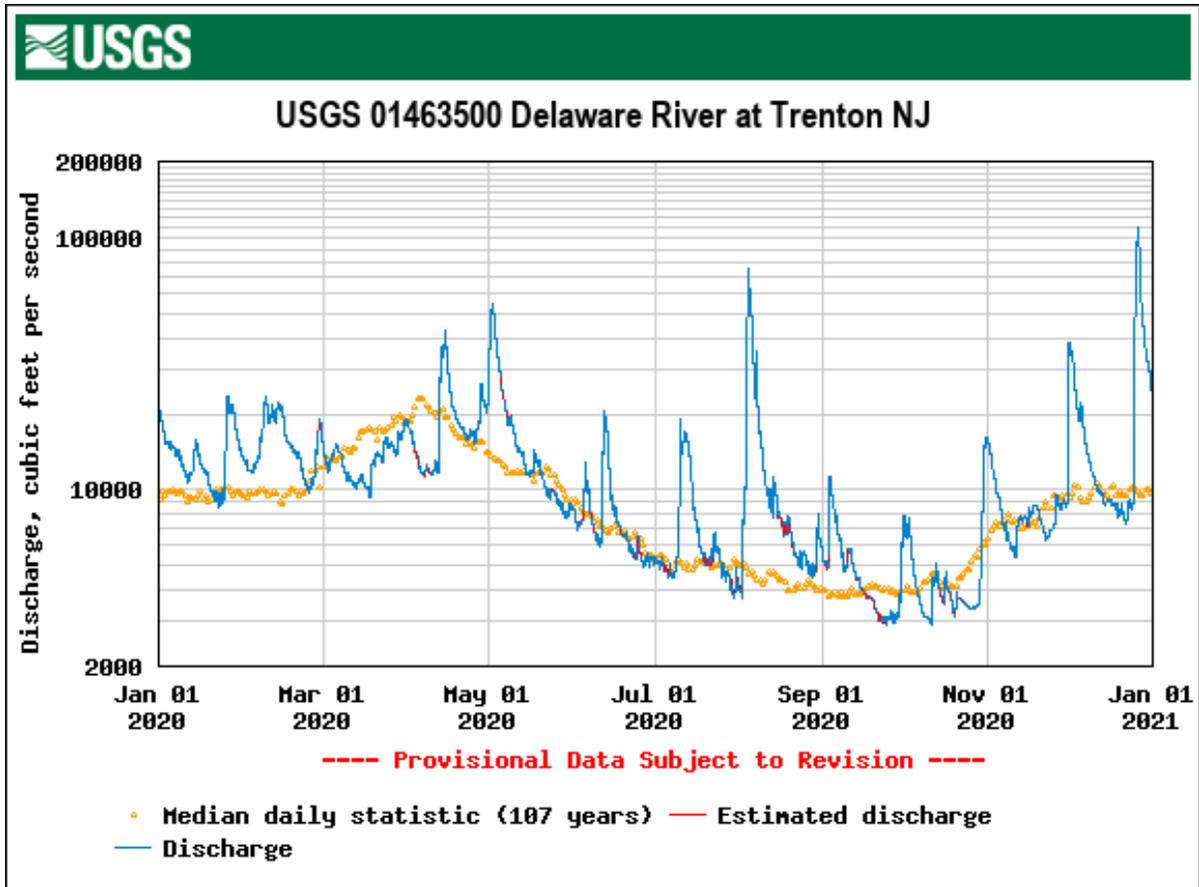


Figure 2. USGS discharge from Delaware River at Trenton (USGS station 01463500) during 2020. Freshwater inflow was generally near long-term median values but was punctuated by several pulses of short duration throughout the year.

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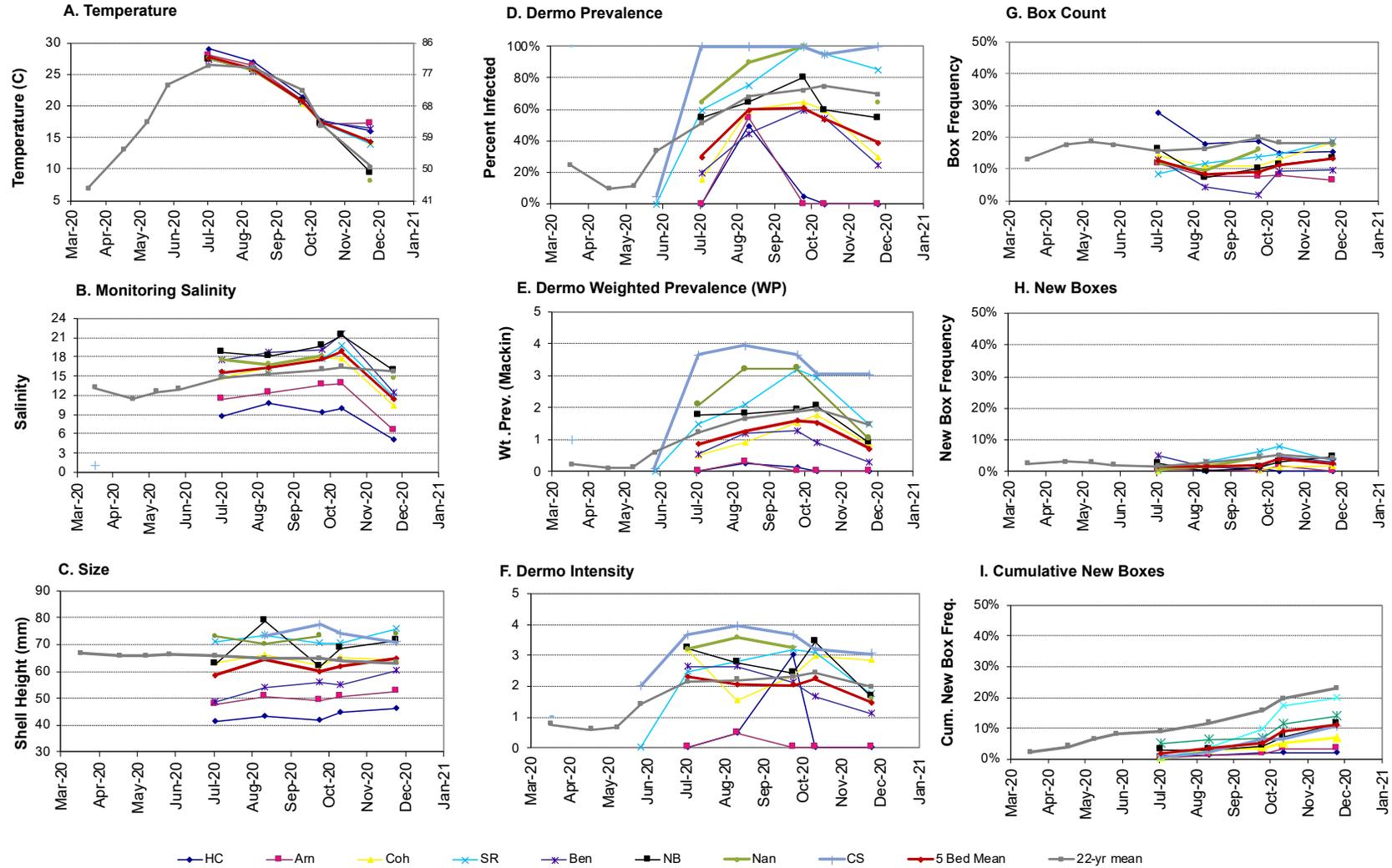


Figure 3. Results of 2020 Seed Bed Monitoring Program. Panels present data as labeled. HC = Hope Creek, Arn = Arnolds, Coh = Cohansey, SR = Shell Rock, Ben = Bennies, NB = New Beds, Nan = Nantuxent, CS = Cape Shore.

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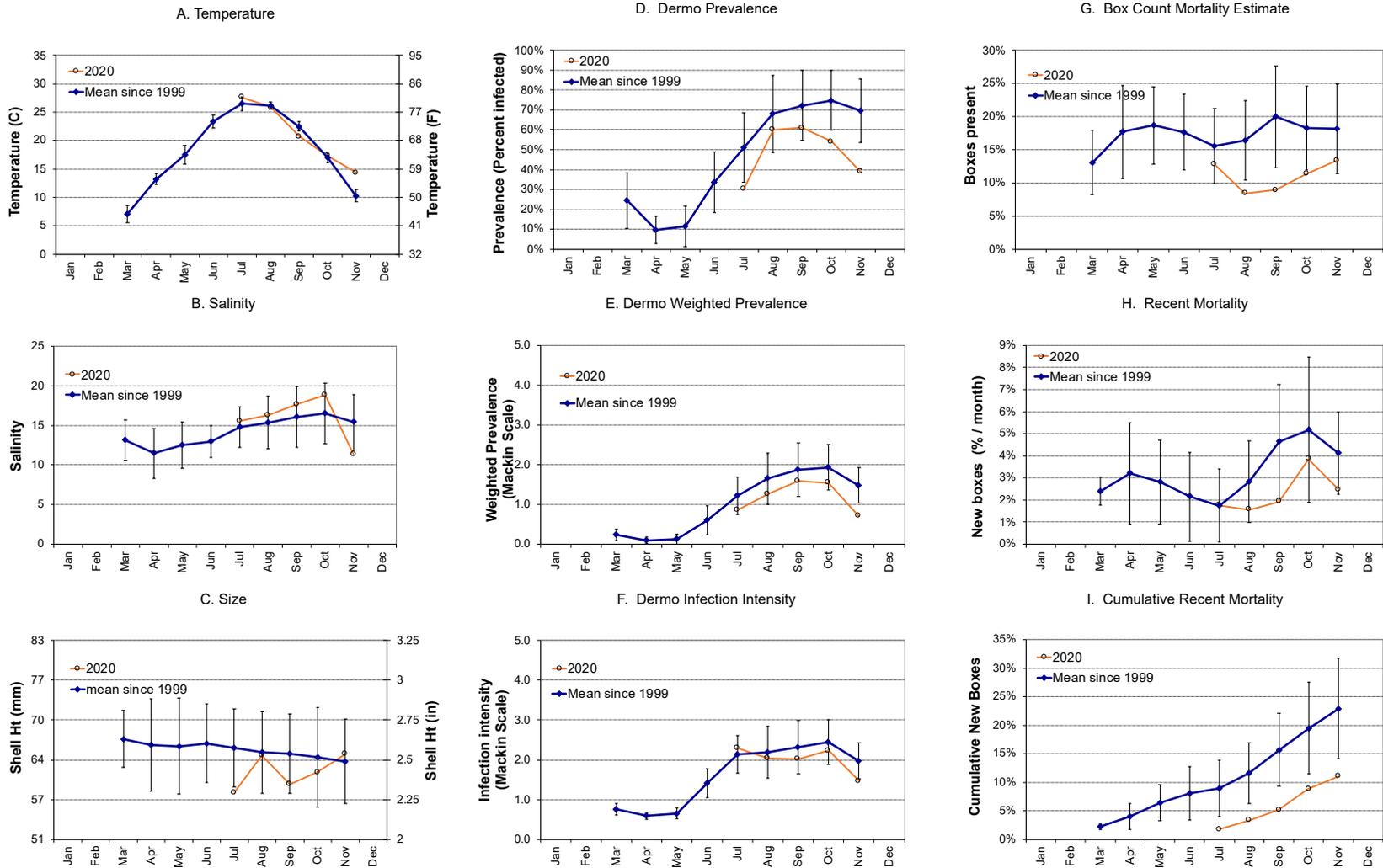


Figure 4. Seasonal patterns of the 2020 average (Arnolds, Cohansey, Shell Rock, Bennies and New Beds) compared to the long-term values. Panels arranged as in Figure 3. Error bars represent one standard deviation.

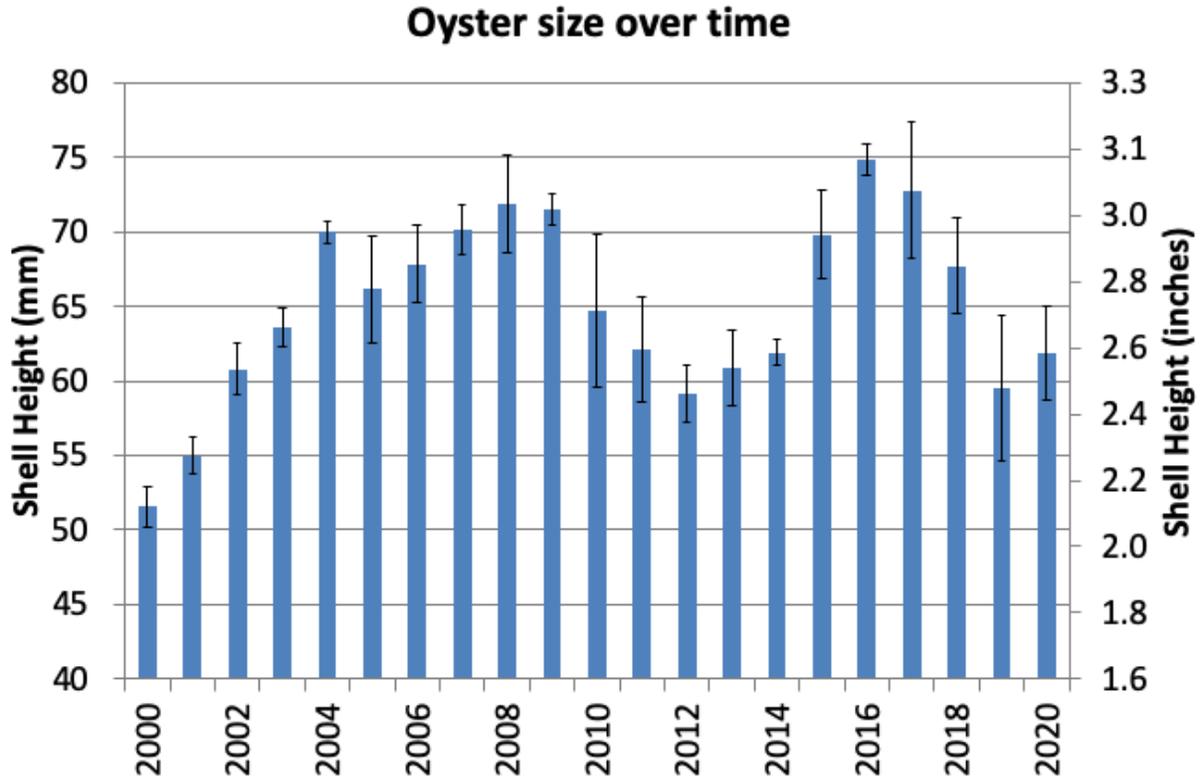


Figure 5. Interannual variation in mean shell height of oysters collected monthly between from Delaware Bay NJ oyster seedbeds. Error bars represent one standard deviation of the mean of all oysters measured throughout each year. N = 50-100 oysters per month from each of the five primary long-term beds (Arnolds, Cohansey, Shell Rock, Bennies and New Beds) sampled from March to November. Samples from 2020 were collected from July to November.

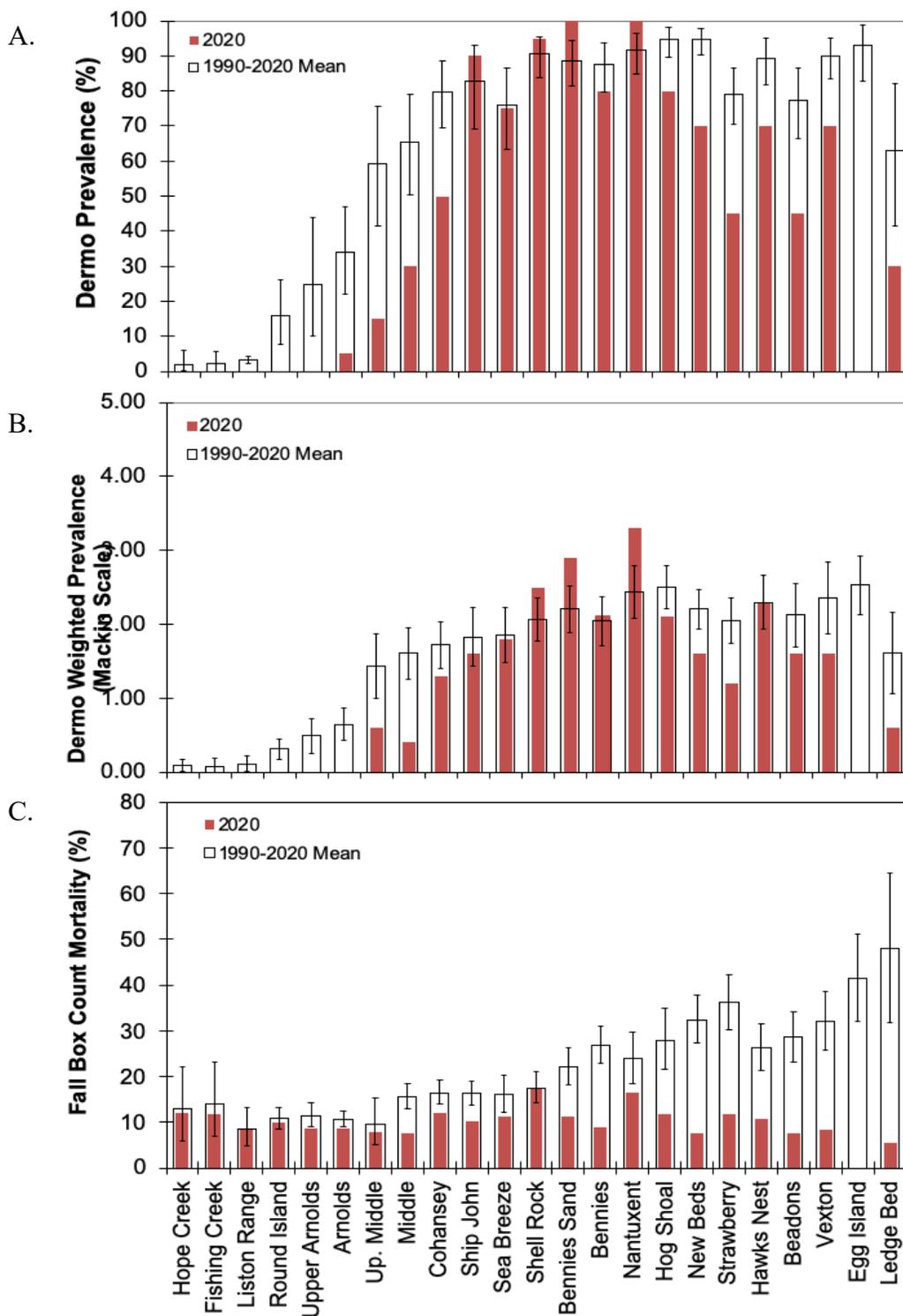


Figure 6. Long-term spatial patterns of dermo prevalence (A), dermo weighted prevalence (B) and natural mortality (C) across the oyster beds. From left to right, beds are listed upbay to downbay. Not all beds have been sampled every year (see Table 3). Egg Island was not sampled in 2020. Error bars represent 95% confidence intervals.

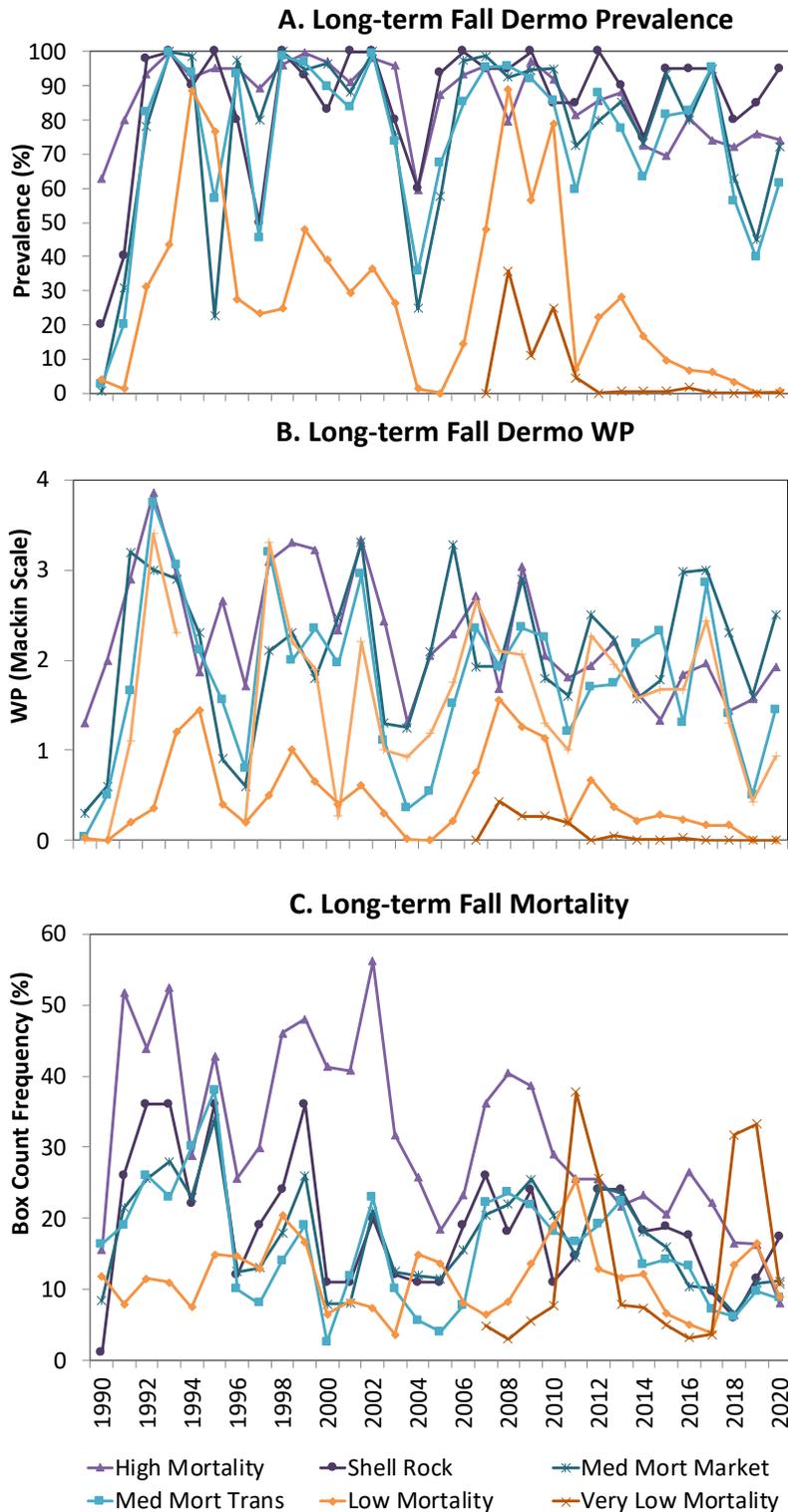


Figure 7 Annual Fall dermo prevalence (A), weighted prevalence (B) and box count mortality (C) on New Jersey Delaware Bay seedbeds. Regions correspond to management regions in Figure 1.

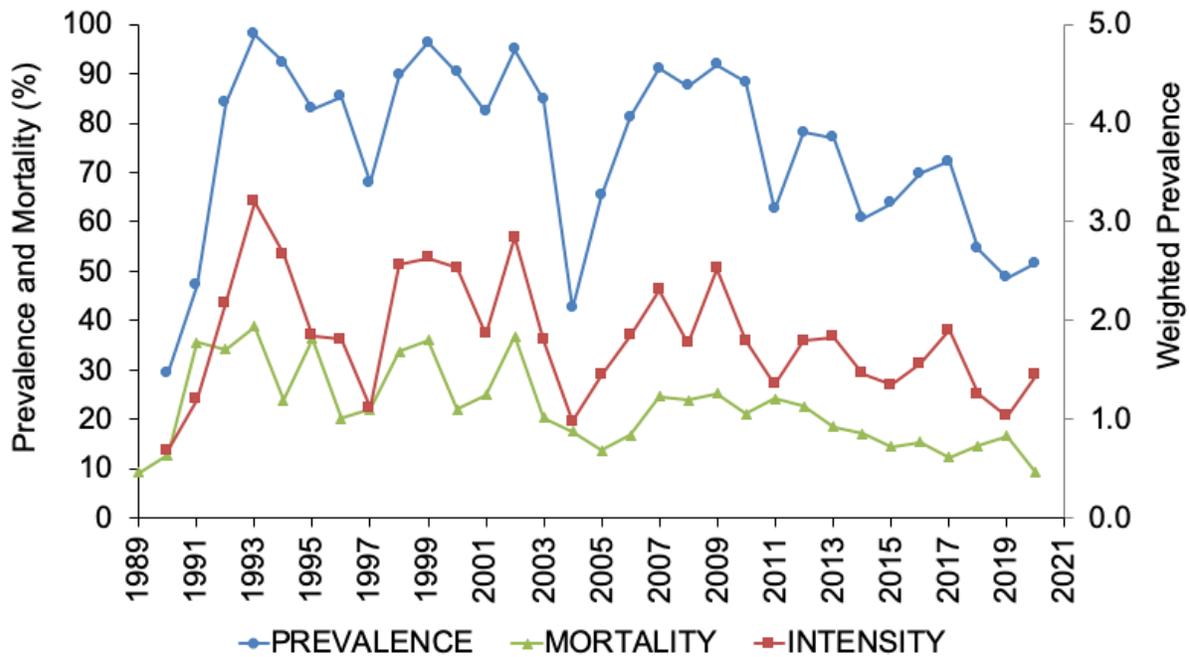


Figure 8. Long-term patterns of Fall dermo prevalence, intensity (weighted prevalence) and mortality averaged across the five beds monitored since 1990 (Arnolds, Cohansey, Shell Rock, Bennies and New Beds). These data appear to show cycles with an approximate periodicity of seven years, and a dampening of the cycling resulting in lower levels of each metric over time.

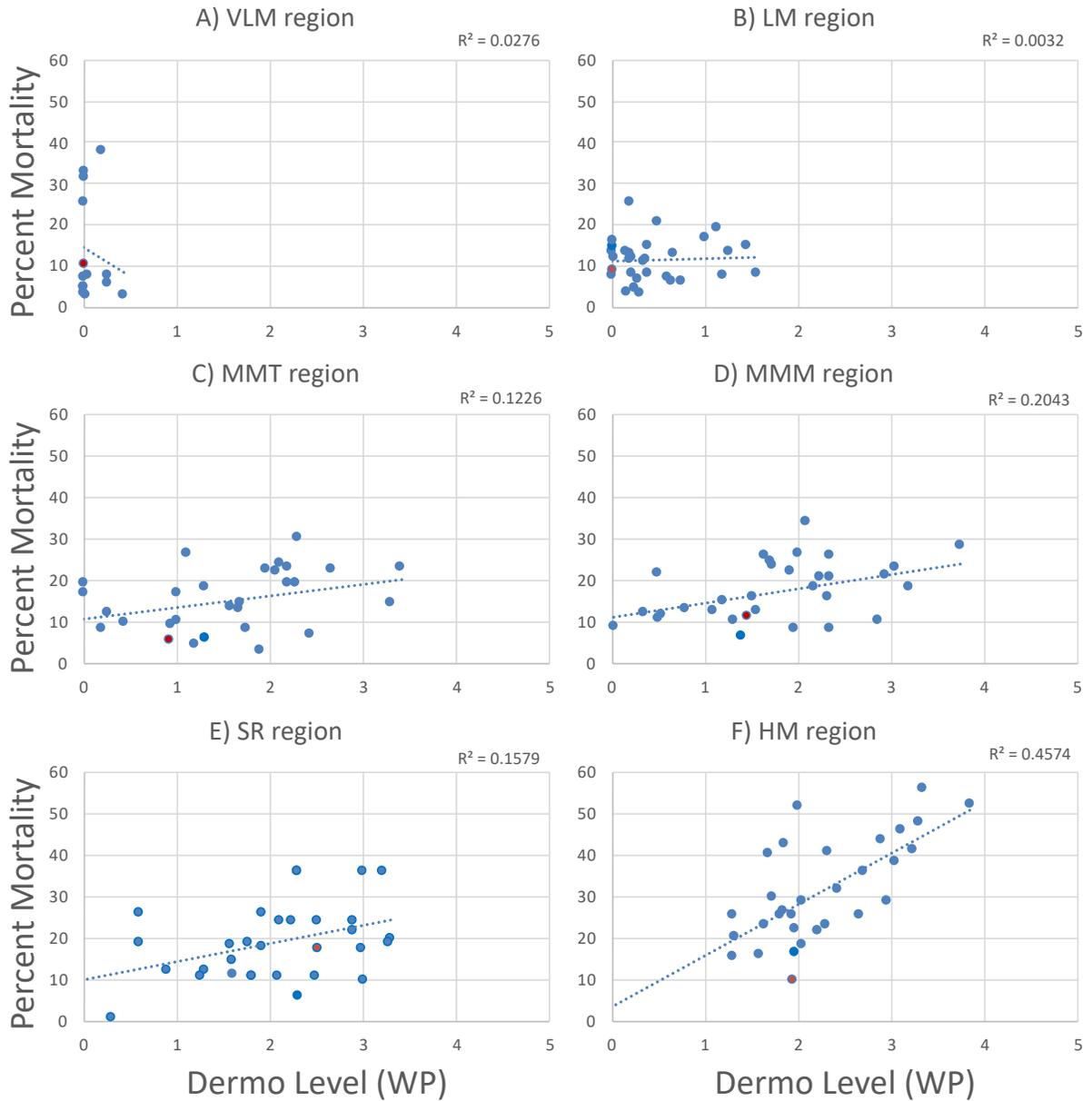


Figure 9. Region mortality as a function of dermo disease levels since 1990 (2007 for the VLM region). Red points represent 2020 data. VLM = Very Low Mortality region, LM = Low Mortality region, MMT = Medium Mortality Transplant region, MMM = Medium Mortality Market region, SR = Shell Rock, and HM = High Mortality Region.

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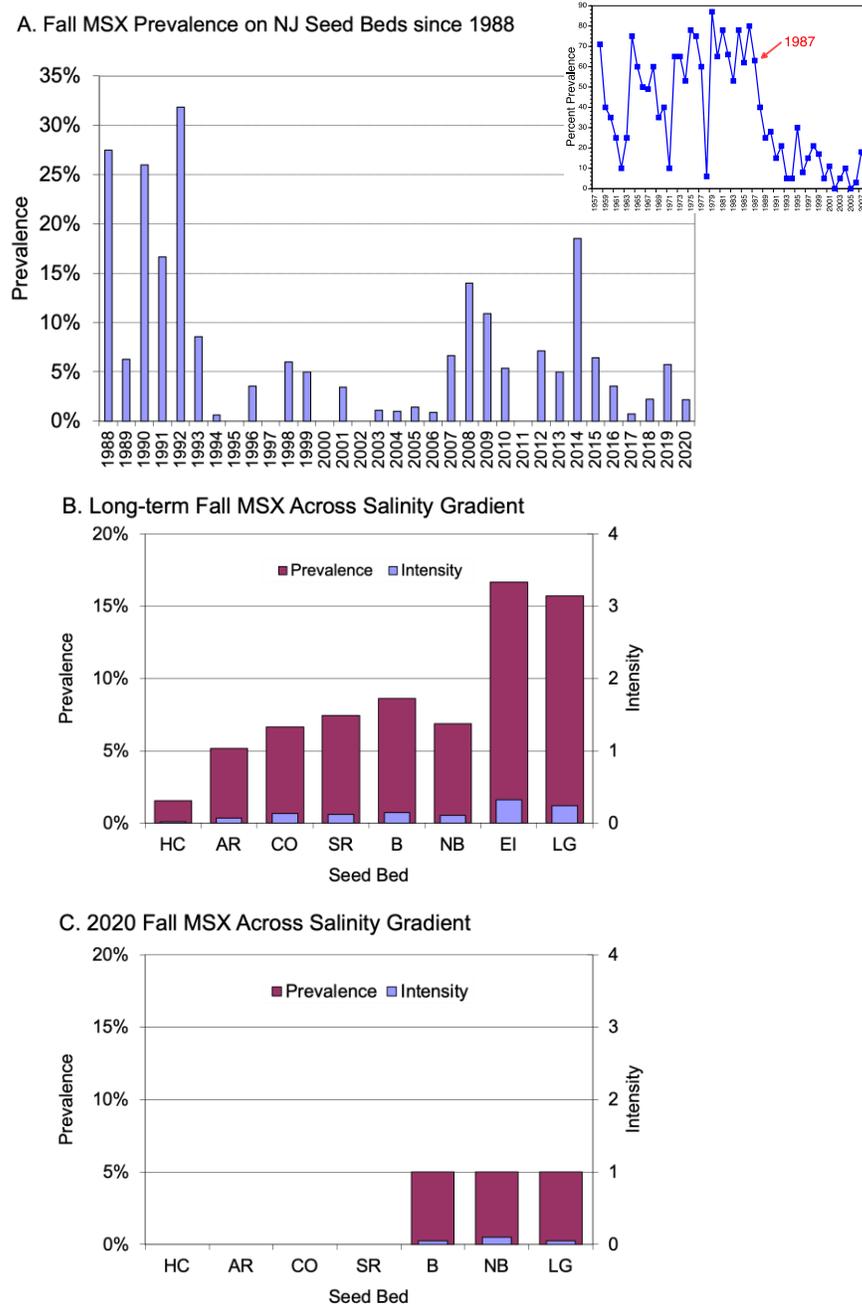


Figure 10. MSX disease on the New Jersey Delaware Bay oyster seedbeds. A. Annual Fall MSX prevalence across all beds since 1988 (2007 for HC). Inset shows lower Delaware Bay levels for comparison from Ford and Bushek (2012). B. Total fall MSX prevalence and intensity (weighted prevalence on a scale of 0 to 4) across seedbed salinity gradient since 1988. C. 2020 Fall MSX prevalence and intensity across seedbeds. HC = Hope Creek, AR = Arnolds, CO = Cohansey, SR = Shell Rock, B = Bennies, NB = New Beds, LG = Ledge.